



Perfluoroalkyl substances exposure and thyroid hormones in humans: epidemiological observations and implications

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Thyroid hormones play crucial roles in normal neurodevelopment of fetus and child. Many chemicals can affect control and homeostasis of thyroid hormones, and eventually lead to various adverse health effects including neurodevelopmental disorders. Perfluoroalkyl substances (PFASs) are among the thyroid disrupting chemicals that can be encountered among general human population. Due to their unique physicochemical characteristics, PFASs have been used as surfactants and surface coating materials in many applications. Therefore, PFASs have been frequently detected in humans and environment worldwide. In cross-sectional studies using nationally representative general human populations of United States, several PFASs have shown significant associations with thyroid hormones. Moreover, among pregnant women and their infants, not only major PFASs such as perfluorooctane sulfonic acid and perfluorooctanoic acid, but also those with shorter or longer carbon chains showed significant associations with thyroid hormones. Often demographic characteristics such as sex, age, and disease status appear to influence the associations between PFASs exposure and thyroid hormones. In general, major PFASs showed hypothyroidism effects among pregnant women and infants. As 8 carbon based PFASs have been phased out, those with shorter or longer carbon chains have been used in growing amount as replacement. However, only limited information is available for their occurrences and toxicity among humans. Further investigations on these substituting PFASs are required. In addition, efforts are warranted to identify sources of and mitigate exposure to these thyroid disrupting chemicals especially during pregnancy and early stages of life.

Keywords: Perfluoroalkyl substances, Biomonitoring, Thyroid, Disruption, population, Pregnant women

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Introduction

Neurodevelopmental disorders among childhood have increased significantly over the past several decades worldwide. In 2014, Centers for Disease Control and Prevention of United States estimated that one out of 68 children show autism spectrum disorder (ASD)¹. This number is very high, compared to the old statistics, e.g., 1 out of 150 children in 2007, and 1 out of 2,500 in 1985². In addition, up to 20% of children, depending on country, exhibit attention deficiency hyperactivity disorders³. Aggressive diagnosis and reporting may explain this phenomenal upsurge of the neurodevelopmental disorders to certain extent. However, importance of environmental causes gains more recognition as determinants of these diseases among the scientists. For example, environmental contaminants are suggested as major causes of ASD, along with diagnostic accretion, greater awareness, and age of parents².

Thyroid hormones are involved in regulation of various processes over early brain development⁴⁾. These hormones are involved in dendritic and axonal growth, neurogenesis, and myelination processes in developing fetus during pregnancy. After birth as well, these hormones are important in myelination of neurons. Therefore, even subclinical changes of thyroid hormones during these susceptible periods of life may lead to delays in neuropsychological development of children⁵⁾.

Numerous observational studies support the importance of thyroid hormones in neurodevelopment of children. Infants with low thyroxine (T4) showed modestly increase risk of ASD in large scale population study in California United States⁶⁾. According to the observations of a prospective cohort study on Dutch mother-child pairs (n=3,839), when the levels free triiodothyronine (T3) of mother were too high or low, the average intelligence quotient (IQ) of her child was lower by 1.4–3.8⁷⁾. As fetal thyroid gland is not sufficiently developed until second trimester of pregnancy, development of fetus will be likely influenced by maternal thyroid hormones. Therefore, the factors that will affect not only thyroid hormones of the

infants but also those of pregnant mothers can affect the IQ and neurodevelopment of the children.

Thyroid disrupting chemicals (TDCs) are suggested to be responsible for, to certain extent, rapid increase of neurodevelopmental diseases worldwide. Numerous chemicals are reported to disrupt thyroid hormone balances. These chemicals include phthalates and phenolic compounds that are widely used in daily life, and persistent organic compounds such as organochlorine pesticides (OCPs), polychlorinated biphenyls (PCBs), polybrominated diphenylethers (PBDEs), and perfluoroalkyl substances (PFASs)⁸⁾.

Perfluoroalkyl substances

PFASs are reported to disrupt thyroid hormone balances by increasing metabolic excretion and inhibiting synthesis of the hormones⁹⁾. Unlike other persistent organic compounds which preferentially partition into lipid, PFASs are unique as these compounds can resist both water and oil. This characteristic is due to their structure with perfluorinated carbon tail and

Table 1. Perfluoroalkyl substances - classification and physicochemical characteristics

| Compound | CAS RN | Molecular weight (g/mol) | log Kow ^{b)} | Serum elimination half-lives | |
|--|------------|--------------------------|-----------------------|--|---|
| | | | | Male | Female |
| Perfluoroalkyl carboxylic acids (PFCAs) | | | | | |
| Short chain PFCAs (CnF _{2n+1} COOH, n<7, PFCAs) ^{a)} | | | | | |
| Perfluorobutanoic acid (PFBA) | 375-22-4 | 214.04 | 3.391±0.597 | *68±35 hr ^{c)} †6 hr (IV) ^{c)} , 9 hr (Oral) ^{c)} | *87±31 hr ^{c)} †1 hr (IV) ^{c)} , 2 hr (Oral) ^{c)} |
| Perfluoropentanoic acid (PFPeA) | 2706-90-3 | 264.05 | 4.391±0.657 | - | - |
| Perfluorohexanoic acid (PFHxA) | 307-24-4 | 314.05 | 4.985±0.710 | *<28 days ^{d)} , †1 hr (IV) ^{g)} | †0.4 hr (IV) ^{g)} |
| Perfluoroheptanoic acid (PFHpA) | 375-85-9 | 364.06 | 5.714±0.759 | †0.10 days (IV) ^{h)} | †0.05 days (IV) ^{h)} |
| Long chain PFCAs (CnF _{2n+1} COOH, n≥7, PFCAs) | | | | | |
| Perfluorooctanoic acid (PFOA) | 335-67-1 | 414.07 | 6.444±0.806 | *3.8±1.7 yr ^{e)} †5.63 days (IV) ^{h)} | *3.3 yr ^{e)} †0.08 days (IV) ^{h)} |
| Perfluorononanoic acid (PFNA) | 375-95-1 | 464.08 | 7.174±0.844 | †29.5 days (IV) ^{h)} , 30.6 days (Oral) ^{l)} | †2.44 days (IV) ^{h)} , 1.4 days (Oral) ^{l)} |
| Perfluorodecanoic acid (PFDA) | 335-76-2 | 514.08 | 7.904±0.860 | †39.9 days (IV) ^{h)} | †58.6 days (IV) ^{h)} |
| Perfluorododecanoic acid (PFDoDA) | 307-55-1 | 614.1 | 9.363±0.888 | - | - |
| Perfluorotridecanoic acid (PFTTrDA) | 72629-94-8 | 664.11 | 10.093±0.901 | - | - |
| Perfluorotetradecanoic acid (PFTeDA) | 376-06-7 | 714.11 | 10.823±0.914 | - | - |
| Perfluoroalkyl sulfonic acids (PFSAs) | | | | | |
| Short chain PFSAs (CnF _{2n+1} SO ₃ H, n<6, PFSAs) | | | | | |
| Perfluorobutane sulfonic acid (PFBS) | 375-73-5 | 300.1 | 1.689±0.719 | *24±7 days ^{f)} †5 hr (IV) ^{f)} , 5 hr (Oral) ^{f)} | *46 days ^{f)} †4 hr (IV) ^{f)} , 8 hr (Oral) ^{f)} |
| Long chain PFSAs (CnF _{2n+1} SO ₃ H, n≥6, PFSAs) | | | | | |
| Perfluorohexane sulfonic acid (PFHxS) | 355-46-4 | 400.11 | 3.053±0.814 | *8.2±5.1 yr ^{e)} †29 days (IV) ^{l)} | *12.8±0.6 yr ^{e)} †2 days (IV) ^{l)} |
| Perfluorooctane sulfonic acid (PFOS) | 1763-23-1 | 500.13 | 4.512±0.862 | *5.4±3.7 yr ^{e)} †38 days (Oral) ^{k)} , 41 day (Oral) ^{k)} | *5.9±1 yr ^{e)} †62 days (Oral) ^{k)} , 71 days (Oral) ^{k)} |
| Perfluorodecane sulfonic acid (PFDS) | 335-77-3 | 600.14 | 5.972±0.891 | - | - |

CAS RN, chemical abstracts service registry number.

*Serum elimination half-lives in human. †Serum elimination half-lives in rat. ^{a)}Buck et al. (2011)¹¹⁾. ^{b)}At 25°C, calculated using Advanced Chemistry Development (ACD/Labs) Software V11.02 (© 1994–2016 ACD/Labs). ^{c)}Chang et al. (2008)¹³⁾. ^{d)}Nilsson et al. (2010)¹⁴⁾. ^{e)}Olsen et al. (2007)¹⁵⁾. ^{f)}Olsen et al. (2009)¹⁶⁾. ^{g)}Chengelis et al. (2009)¹⁷⁾. ^{h)}Ohmori et al. (2003)¹⁸⁾. ⁱ⁾Tatum-Gibbs et al. (2011)¹⁹⁾. ^{j)}Sundström et al. (2012)²⁰⁾. ^{k)}Chang et al. (2012)²¹⁾.

polar head¹⁰). For this property, PFASs have been widely used as surfactants and surface coatings in various applications in industry and commerce since 1950s, such as textiles, food contact paper, cookware, carpets, cosmetics, photographic emulsifiers, lubricants, paints, fire-fighting foams, and food packaging^{11,12}.

PFASs are highly fluorinated synthetic aliphatic substances of which H substituents are replaced by F (C_nF_{2n+1}). Major PFASs are shown in Table 1^{11,13-21} and Fig. 1. By the component, PFASs are grouped into perfluoroalkyl sulfonic acids (PFSAs) and perfluoroalkyl carboxylic acids (PFCAs). Generally longer carbon chain PFASs, e.g., long-chain PFSAs (number of C ≥ 6) and PFCAs (number of C ≥ 7), are reported to be more bioaccumulative and persistent compared to the short-chain PFASs¹¹. Definition of long-chain is different for PFCAs and PFSAs, as PFSA with 6 carbons (perfluorohexane sulfonic acid [PFHxS]) is more bioaccumulative than perfluorohexanoic acid (PFHxA).

Recently, heavily used long-chain PFASs such as perfluorooctane sulfonic acid (PFOS) have been phased out, and shorter carbon chain PFASs have been used as replacements in increasing amount²². Therefore, the occurrence of the substitutes in both environment and biota has increased worldwide. In breastmilk samples of Korean lactating women (n=265), while perfluorooctanoic acid (PFOA) and PFOS were detected in 98.5% of the samples, short carbon chain PFASs such as perfluoropentanoic acid (PFPeA), PFHxA, and perfluoroheptanoic acid, were detected in high frequency ranging between 67.4% and 81.8%. The levels of PFPeA and PFHxA in breastmilk were the greatest compared to the previous reports suggesting that these compounds that replace the long chain PFASs are used in increasing amount over time²³. Longer carbon chain PFASs (number of C ≥ 9) are used also in

growing amount. Reflecting this trend, in Australia, the serum levels of PFOS and PFOA showed decreasing trend between 2002 and 2013, but perfluorododecanoic acid (PFDoDA) showed an increase²⁴.

Biological half-lives of PFASs

Because of the different composition and chain length, biological half-lives of PFASs vary by structure. Long-chain compounds like PFHxS, PFOS, and PFOA have longer half-lives than the PFASs with shorter carbon chains¹⁶. For example, half-life of serum PFOA is estimated up to 8.5 years, but those for perfluorobutanoic acid and PFHxA are ranging between 3 and 28 days^{13,14}. Biological half-lives of other short-chain PFASs are also expected to be relatively short.

PFASs exposure and thyroid hormone disruption in general populations

Association studies on general population without occupational exposure are accumulating. Several big scale studies employing nationally representative populations are also conducted. In addition, studies on susceptible human populations such as pregnant women and newborns or children are conducted. Reports on nationally representative human populations are mostly based on US National Health and Nutrition Examination Survey (NHANES). Among general US population, PFOS, PFOA, PFHxS, and perfluorononanoic acid (PFNA) are most frequently measured and reported. The exposure to these chemicals showed significant association with thyroid hormone levels (Table 2)²⁵⁻²⁹.

Among adults with current thyroid diseases, PFOA or PFOS were detected at relatively higher levels²⁵. While many PFASs

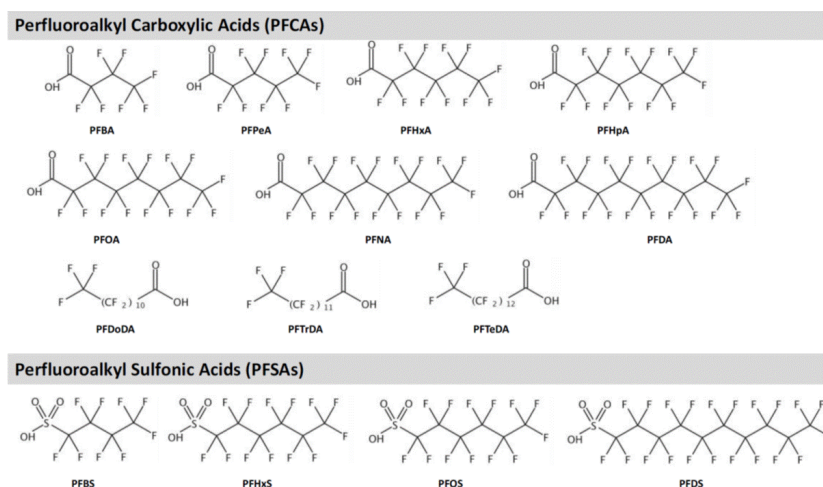


Fig. 1. Structures of major perfluoroalkyl substances. PFBA, perfluorobutanoic acid; PFPeA, perfluoropentanoic acid; PFHxS, perfluorohexane sulfonic acid; PFHpA, perfluoroheptanoic acid; PFOA, perfluorooctanoic acid; PFNA, perfluorononanoic acid; PFDoDA, perfluorododecanoic acid; PFTrDA, perfluorotridecanoic acid; PFTeDA, perfluorotetradecanoic acid; PFBS, perfluorobutane sulfonic acid; PFOS, perfluorooctane sulfonic acid; PFDS, perfluorododecane sulfonic acid.

showed significant associations with thyroid hormone levels, the directions of association were different by sex²⁶, and age²⁷. According to the survey 2007–2010, the PFOA and PFHxS levels showed positive associations with total thyroid hormones among female (n=509), however, among male population (n=672), negative association was observed between PFHxS and free T4 (ft4) levels²⁶. Among the adolescents in the 2011–2012 survey, males (n=158) showed positive associations between PFOS or PFNA and thyroid stimulating hormone (TSH), while females (n=145) showed a negative association between PFOA and TSH²⁷. In the same survey round (NHANES, 2011–2012), effects of age were also observed. Female subjects of 20 to 40 years of age (n=268) showed positive associations between study PFASs, i.e., PFOA, PFHxS, PFNA, or PFOS, and TSH. However, among females of 60 to 80 years of age, a negative association between PFHxS and TSH was observed²⁷. By levels of thyroid peroxidase (TPO) antibody (Ab) and urinary iodine, the association between PFASs and thyroid hormones differs²⁸. In NHANES 2007–2008, subjects with normal levels of TPOAb and urinary iodine (n=1,012) exhibited a positive association between PFOA and free T3. However, among those who showed abnormal range of these markers (n=26) negative associations were observed between PFHxS or PFOA and ft4²⁸. While it was not nationally representative sampling, among general

Korean populations (>12 years of age, n=644), Ji et al.³⁰ reported a negative association of perfluorotridecanoic acid with tT4, and a positive association with TSH, in a small industrial city of Siheung, Korea.

The influences of sex and age on the reported associations may be due to the biological half-lives of PFASs. The biological half-lives of major PFASs tend to be shorter among females, probably because of menstruation³¹. The fact that post-menopausal women, e.g., women of 60–80 years of age, showed different direction of association compared to menstruating women²⁷, may also support this hypothesis. In addition, the interaction between sex hormones and thyroid hormones cannot be excluded. Several human thyroid disorders are accompanied by changes in sex steroids³². In other animals, e.g., rainbow trout, estradiol exposure decreased activities of T4-outer ring deiodination (ORD) and rT3-ORD in liver, but increased activity of T3-inner ring deiodination in kidney in both sexes³³. In addition, estradiol treatment in immature rainbow trout, caused decrease of T3 in plasma³⁴.

As shown above, among general populations with various demographic characteristics, such as sex, age, and disease status, it is not easy to observe consistent directions of association. Moreover, potential effects of other TDCs that may present in the study populations cannot be ignored. Since the exposure

Table 2. Associations between exposure to PFASs and thyroid hormones among nationally representative populations of United States

| PFASs | Population age (n) | Study period | Thyroid measure | Association | Reference |
|--|----------------------------|--------------|-------------------------|---|-------------------------------------|
| PFOA PFOS | General population (3,974) | 1999–2006 | Current thyroid disease | -PFOA highest quartile shows odd ratio (OR) for current thyroid disease of 2.24 (95% CI, 1.38–3.65) in females, and 2.12 (0.93–4.82) for male -PFOS highest quartile shows OR for current thyroid disease of 2.68 (1.03–6.98) in males | Melzer et al. (2010) ²⁵ |
| PFOS, PFOA, PFDeA, PFHxS, PFNA | >12 years (1,832) | 2007–2008 | ft3, ft4, tT3, tT4, TSH | -PFOA in positive associations with TSH and tT3 -PFHxS in positive association with tT4 | Jain (2013) ²⁹ |
| PFOA, PFHxS, PFNA, PFOS | >18 years (1,525) | 2007–2008 | ft3, ft4, tT3, tT4, TSH | -Among population with normal range of TPOAb and urinary iodine (n=1,012), PFOA in positive association with ft3 -Among population with abnormal range of TPOAb and urinary iodine (n=26), PFHxS and PFOA in negative associations with ft4. All 4 PFASs in positive associations with ft3, tT3, ft3/ft4, or TSH | Webster et al. (2016) ²⁸ |
| PFOA, PFHxS, PFNA, PFOS | ≥20 years (1,181) | 2007–2010 | ft3, ft4, tT3, tT4, TSH | -Among female (n=509), PFOA in positive association with tT3; PFHxS in positive associations with tT4 and tT3 -Among male (n=672), PFHxS in negative association with ft4 | Wen et al. (2013) ²⁶ |
| PFOA, PFHxS, PFNA, PFOS | ≥12 years (1,682) | 2011–2012 | ft3, ft4, tT3, tT4, TSH | -No associations observed in whole population -Among females of 20–40 years of age (n=268), all PFASs in positive associations with TSH. But in females of 60–80 years of age, PFHxS in negative association with TSH -Among female juveniles (n=145), PFOA in negative association with TSH | Lewis et al. (2015) ²⁷ |

PFAS, perfluoroalkyl substance; PFOA, perfluorooctanoic acid; PFOS, perfluorooctane sulfonic acid; PFDeA, perfluorodecanoic acid; PFHxS, perfluorohexane sulfonic acid; PFNA, perfluorononanoic acid; TSH, thyroid stimulating hormone; ft3, free triiodothyronine; ft4, free thyroxine.

levels of PFASs are lower compared to experimental studies or occupational settings, their thyroid disrupting effects could be compensated by the effects of other chemicals, if present.

PFASs exposure and thyroid hormone disruption in susceptible human populations

Many cross-sectional and often cohort studies have been conducted on susceptible populations including pregnant women and infants (Table 3). These studies reflect the importance of thyroid hormones that play crucial roles in developing fetus and infants. Unlike big observation studies employing nationally representative populations, these studies on susceptible groups are smaller in population size, but often measured more diverse types of PFASs.

Generally, exposure levels of PFASs showed negative associations with thyroid hormones but positive association with TSH, although deviations were often noted. Chemicals, Health and Pregnancy study conducted among pregnant women of Vancouver Canada (n=152) reported a positive association between PFNA and TSH. Among TPOAb abnormal group, PFASs showed positive associations with TSH, and negative associations with fT4³⁵. In a mother and child cohort of Norway, PFASs such as perfluorodecanoic acid (PFDA) and perfluoroundecanoic acid (PFUnDA) showed also negative associations with T3, and PFOS showed a positive association with TSH³⁶. Similarly, in Norwegian pregnant women (n=930), PFOS that was measured at 18 weeks of pregnancy showed a positive association with TSH³⁷.

Hypothyroidism by PFASs exposure has also been reported in Asian countries. Among Taiwanese mothers of late pregnancy (n=285), maternal PFHxS levels showed a positive association with TSH, and PFASs such as PFNA, PFUnDA, or PFDoDA, showed negative associations with free and total T4³⁸. In addition, maternal PFASs showed negative associations with cord total T3 and T4 levels, and maternal PFDA showed a negative association with cord total T3 levels. In Chinese mother and child pairs recruited in Beijing (n=157), maternal perfluorododecanoic acid also showed negative associations with free and total thyroid hormones. In addition, the levels of long chain PFASs, e.g., PFOA, PFNA, PFDA, PFUnDA, PFDoDA, PFHxS, or PFOS in cord serum, showed negative associations with maternal fT3 levels³⁹.

However, in the same population of China, many long chain PFASs measured in maternal serum showed negative associations with maternal TSH, not supporting hypothyroidism effects of PFASs exposure³⁹. Some studies also reported similar inconsistent observations^{22,40,41}. In Hokkaido cohort with matched mother and child pairs (n=39), PFOS levels measured in 24 weeks of pregnancy showed a negative association with maternal TSH, but a positive association with cord serum TSH. In another mother and child pair study (n=83), PFOA showed a positive association with T4 but only among female babies in the Netherlands (n=31)⁴⁰. The association between short chain PFASs exposure and thyroid hormone levels was also reported.

In a retrospective cohort study between 2006 and 2010 in Korea (n=279), a positive association between PFPeA and thyroid hormones was observed, while PFHxS was also found to be positively associated in the same population⁴¹.

Unlike studies on nationally representative population studies, studies on pregnant women and matching infants often measured shorter or longer carbon chain PFASs and reported their associations with thyroid hormones^{39,41,42}. Especially for short carbon chain PFASs, however, limitations of representation of exposure windows should be considered in association studies because of their relatively shorter biological half-lives. Interpretation of cord thyroid hormone levels warrants caution as well, because cord hormone measurement may be influenced by thyroid hormones of maternal origin. Cord thyroid hormone levels can also be changing until shortly after the birth, by several delivery-related factors such as mode and time of delivery.

Proposed mechanisms of thyroid disruption

Most experimental studies and to lesser extent human observational studies support hypothyroidism effects of PFASs exposure, e.g., decrease of thyroid hormones, or increase of TSH⁵. PFASs can inhibit synthesis of thyroid hormones. PFASs are reported to decrease TPO activity in cells⁴⁴. These compounds can also modulate thyroid hormone signaling and function of nuclear hormone receptor. In addition, PFASs can influence metabolism and excretion of thyroid hormones. PFASs can compete with T4 for binding with transthyretin (TTR) which is a transporter of thyroid hormones, in rat. Binding of PFASs with TTR results in increase of fT4 levels in circulation, leading to facilitated metabolic excretion of this hormone^{45,46}. PFOS upregulates T4 metabolic genes in rat, and also enhance hepatic glucuronidation and subsequently excretion of T4⁴⁶.

Implications and conclusion

Human observational studies show that PFASs exposure among general human population is associated with altered thyroid hormone levels. However, it should be noted that observational studies based on general human populations have several intrinsic limitations that should be considered in interpretation of the results. First, one cannot rule out potential effects of other TDCs. Phthalates, bisphenols, PCBs, OCPs, and PBDEs have been suggested as TDCs among general human populations⁴⁷⁻⁴⁹. Exposure levels of PFASs among general populations tend to be low, and therefore, in the presence of other important TDCs that are not measured, the dose-response association can be significantly distorted often leading to wrong representation. Second, temporal comparability should be considered especially in cross-sectional or case-control studies⁵. Since most long-chain PFASs have long biological half-lives, one time measurement of the exposure may reflect relatively long

Table 3. Associations between exposure to PFASs and thyroid hormones among pregnant women or infants

| PFASs | Study population | Study period | Thyroid measure | Association | Reference |
|---|--|--------------|--|--|--|
| PFHxS, PFOA, PFOS, PFNA, PFDeA, PFUnDA, PFDoDA, PFHpA, PFHxA | Taiwan Maternal and Infant Cohort Study (285 mothers at 3rd trimester, and 116 cord blood) | 2000–2001 | ft4, tT4, tT3, TSH | -Among mothers, PFHxS in positive association with TSH; maternal PFNA, PFUnDA, or PFDoDA in negative associations with maternal ft4 and tT4 -Maternal PFNA, PFUnDA, or PFDoDA in negative associations with cord tT4 and tT3; Maternal PFDeA in negative association with cord tT3. | Wang et al. (2014) ³⁸⁾ |
| PFOS, PFOA | Hokkaido Study on the Environment and Children's Health (392 pairs) | 2002–2005 | ft4, TSH | -PFOS in negative association with maternal TSH, but positive association with cord TSH -PFOA shows no association (mother during 24 weeks of gestation and 5 days after delivery) | Kato et al. (2016) ²²⁾ |
| PFOS, PFOA, PFHxS, PFNA, PFUnDA, PFHpS, PFDeA | Northern Norway Mother and Child Cohort Study. Pregnant women (930) | 2003–2004 | TSH | -PFOS in positive association with TSH (18 weeks of gestation) | Wang et al. (2013) ³⁷⁾ |
| PFOS, PFOxS, PFOA | Pregnant women, 2nd trimester (96 cases, 175 controls) Edmonton, Alberta, Canada | 2005–2006 | Hypo-thyroidism | -No association | Chan et al. (2011) ⁴³⁾ |
| PFOA, PFHxS, PFNA, PFOS | Euthyroid pregnant women (152) Vancouver, Canada | 2007–2008 | ft4, tT4, TSH | -PFNA in positive association with TSH -Among TPOAb normal group, no association -Among TPOAb abnormal group (n=14), PFNA, PFOS, or PFOA in positive associations with TSH; all PFASs in negative associations with ft4 | Webster et al. (2014) ³⁵⁾ |
| 26 PFASs | Northern Norway Mother and Child Contaminant Cohort Study. 2nd trimester (441) | 2007–2009 | ft3, ft4, tT3, tT4, TSH, TBG, TTR, Albumin, TBI, TPOAb | -PFOS in positive association with TSH -PFDA or PFUnDA in negative association with T3 | Berg et al. (2015) ³⁶⁾ |
| PFHxS, PFHpS, PFOS, PFOA, PFNA, PFDeA, PFUnDA, PFTrDA | Pregnant women (n=44) and infant (n=43) pair | 2008–2009 | tT3, tT4, TSH | -Maternal PFOS in negative association with cord tT3 -Maternal PFOA in positive association with cord TSH -Maternal PFTrDA in negative associations with cord tT4 and tT3 | Kim et al. (2011) ⁴²⁾ |
| PFOA, PFOS, PFTrDA, PFHxS, PFUnDA, PFNA, PFPeA, PFDeA, PFDoDA, PFTeDA | Korea Ewha Birth & Growth Retrospective Cohort (279) | 2006–2010 | T3, T4, TSH | -Among female infants, PFPeA, or PFHxS in positive association with T3 or T4; PFNA in negative association with TSH -Among all infants, PFPeA in positive association with T4 (PFASs measured in maternal serum of 24–28 weeks of gestations. Thyroid hormones measured in cord blood serum) | Shah-Kulkarni et al. (2016) ⁴¹⁾ |
| PFOA, PFOS, other POPs | Netherlands LINK study (83 pairs) | 2011–2013 | T4 in heel prick blood spots | -Among girls (n=31), PFOA in positive association with T4 -Among boys, no associations | de Cock et al. (2014) ⁴⁰⁾ |
| PFHxS, PFOS, PFOA, PFNA, PFDeA, PFUnA, PFDoA | Beijing, China, Mother-infant pair (157) | 2013 | ft3, ft4, tT3, tT4, TSH | -Maternal PFNA, PFDeA, PFUnDA, PFDoDA, or PFOS in negative associations with maternal TSH; Maternal PFDoA in negative associations with ft3, tT3, ft4, or tT4 -Cord PFOA, PFNA, PFDeA, PFUnDA, PFDoDA, PFHxS, or PFOS in negative associations with maternal tT3; cord PFOS in negative association with maternal TSH | Yang et al. (2016) ³⁹⁾ |

PFAS, perfluoroalkyl substance; PFHxS, perfluorohexane sulfonic acid; PFOA, perfluorooctanoic acid; PFOS, perfluorooctane sulfonic acid; PFNA, perfluorononanoic acid; PFDeA, perfluorodecanoic acid; PFUnDA, perfluoroundecanoic acid; PFDoDA, perfluorododecanoic acid; PFHpA, perfluoroheptanoic acid; PFHxA, perfluorohexanoic acid; PFTrDA, perfluorotridecanoic acid; PFPeA, perfluoropentanoic acid; PFHpS, perfluoroheptane sulfonate; PFTeDA, perfluorotetradecanoic acid; PFUnA, perfluoroundecanoic acid; PFDoA, perfluorododecanoic acid; TSH, thyroid stimulating hormone; ft3, free triiodothyronine; ft4, free thyroxine; TPOAb, thyroid peroxidase antibody .

period of exposure window. However, certain PFASs especially with shorter carbon chains have notably shorter biological half-lives and therefore lack temporal representativeness compared to their longer chain analogues. Moreover, influences of biological factors such as age and sex on effects of PFASs on thyroid hormones should add more complexity in elucidating underlying association.

Various PFASs were found to be associated with thyroid hormone levels among susceptible populations such as pregnant women and infants as well as general population. PFASs may affect synthesis, metabolism and excretion of thyroid hormones. Even within normal range, changes in thyroid hormone levels during sensitive periods of developments may cause permanent damages in intelligence and neurodevelopment. Because of their notable persistence and toxicity, some PFASs, such as PFOS and PFOA have been regulated worldwide. However, there are very limited information on their substituting analogues, e.g., shorter or longer carbon chain PFASs. For these substitutes, toxicity information is lacking and exposure levels have not been characterized. Currently available information based on association studies with such substituting PFASs suggests that these chemicals may also disrupt thyroid among pregnant women and infants. Further investigations including exposure assessment and toxicity studies, along with their associations with thyroid related health effects among susceptible human populations are warranted.

Conflict of interest

No potential conflict of interest relevant to this article was reported.

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